

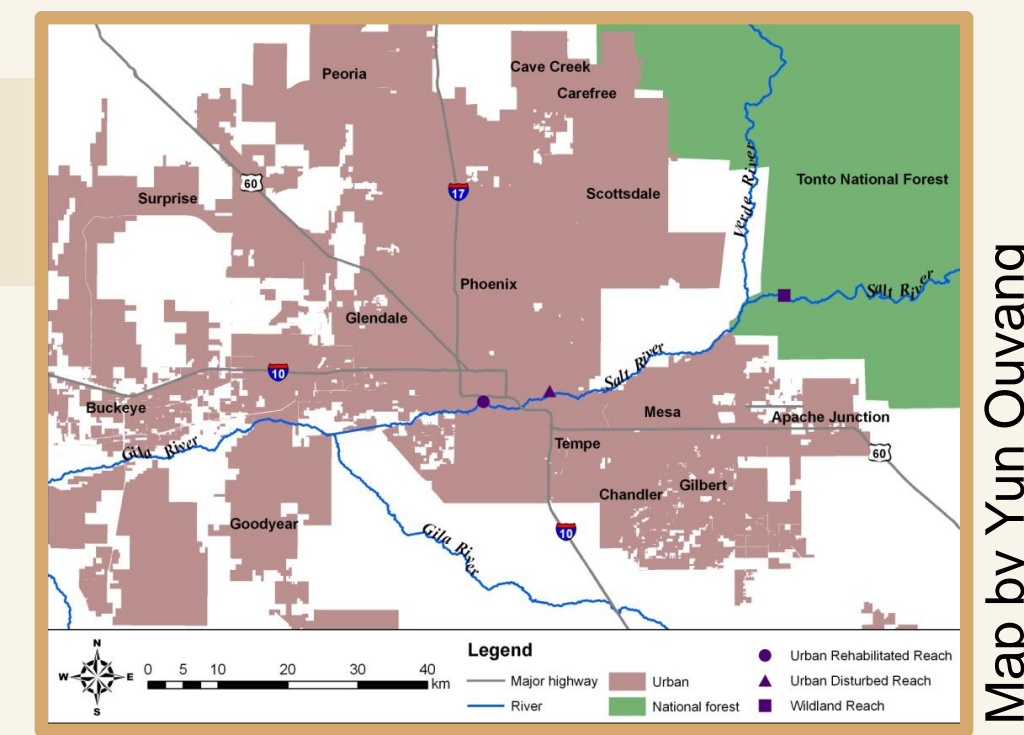
## Background

- Riparian ecosystems support a high diversity and abundance of wildlife species and are used as migration corridors. Unfortunately, 70% of the riparian forests of the lower 48 states in the USA have been converted to other land uses (Turner et al. 1998).
- To mitigate for habitat loss, rehabilitation of degraded land can be performed. However, there is little published information on how rehabilitation activities impact non-avian wildlife communities in riparian ecosystems.
- Herpetofauna are an essential element for healthy ecosystems although they are understudied in riparian communities. Herpetofauna occurrence and abundance are important to monitor because herpetofauna respond to structural changes in their environment.



## Study Sites

Along the Salt River, AZ, we established 24 transects along 3 reaches which vary in terms of urbanization and vegetation.



**Urban Rehabilitated Reach** Phoenix Metropolitan area, recently rehabilitated  
**Urban Disturbed Reach** Phoenix Metropolitan area, highly disturbed  
**Wildland Reach** Tonto National Forest, pristine conditions

## Objectives

- Compare herpetofauna community in terms of abundance, species richness, and diversity among the 3 reaches
- Compare microhabitat characteristics among the 3 reaches
- Develop ecological models to predict occurrence and abundance of herpetofauna (*work in progress*)



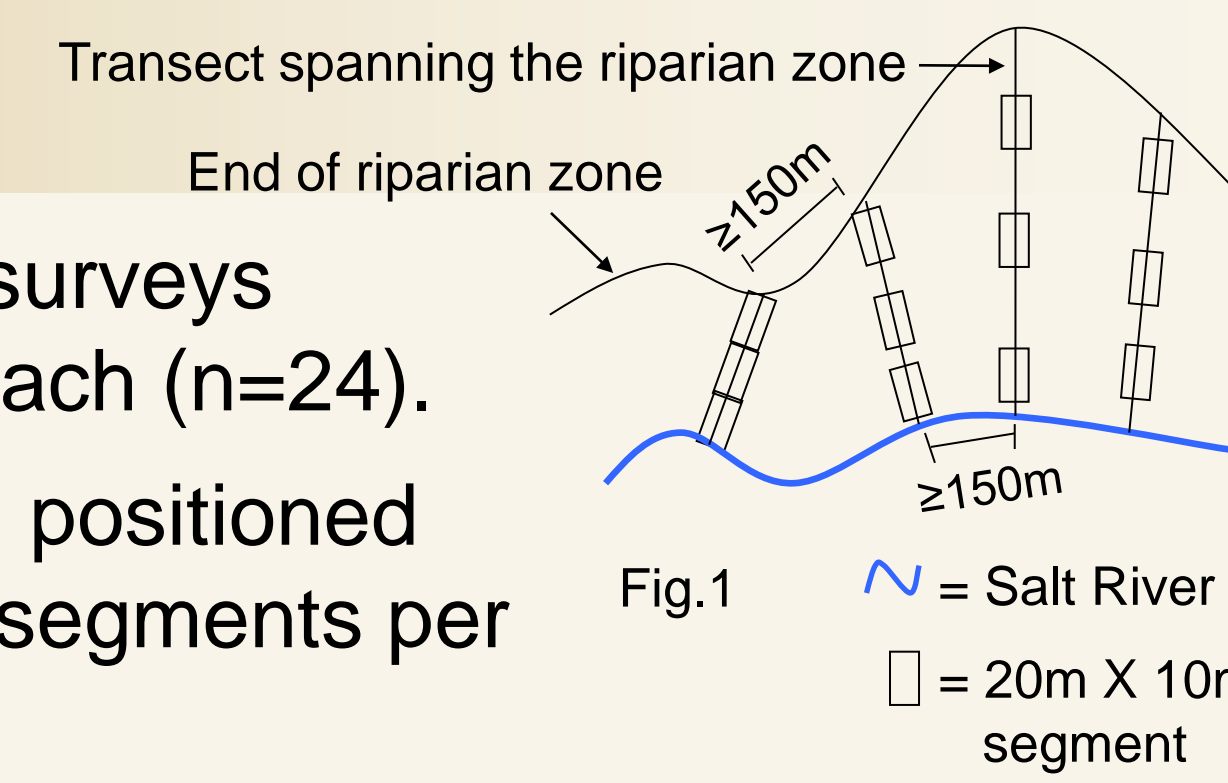
## Acknowledgments and References

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Tilman, D. and J.A. Downing. 1994. Biodiversity and stability in grasslands. *Nature* 367:363-365.  
 Turner, M.G., S.R. Carpenter, E.J. Gustafson, R.J. Naiman, and S.M. Pearson. 1998. Land use. Pages 37-61 in M.J. Mac, P.A. Opler, P. Doran, and C. Haecker, editors. Status and trends of our nation's biological resources. Volume 1. National Biological Service, Washington, DC.

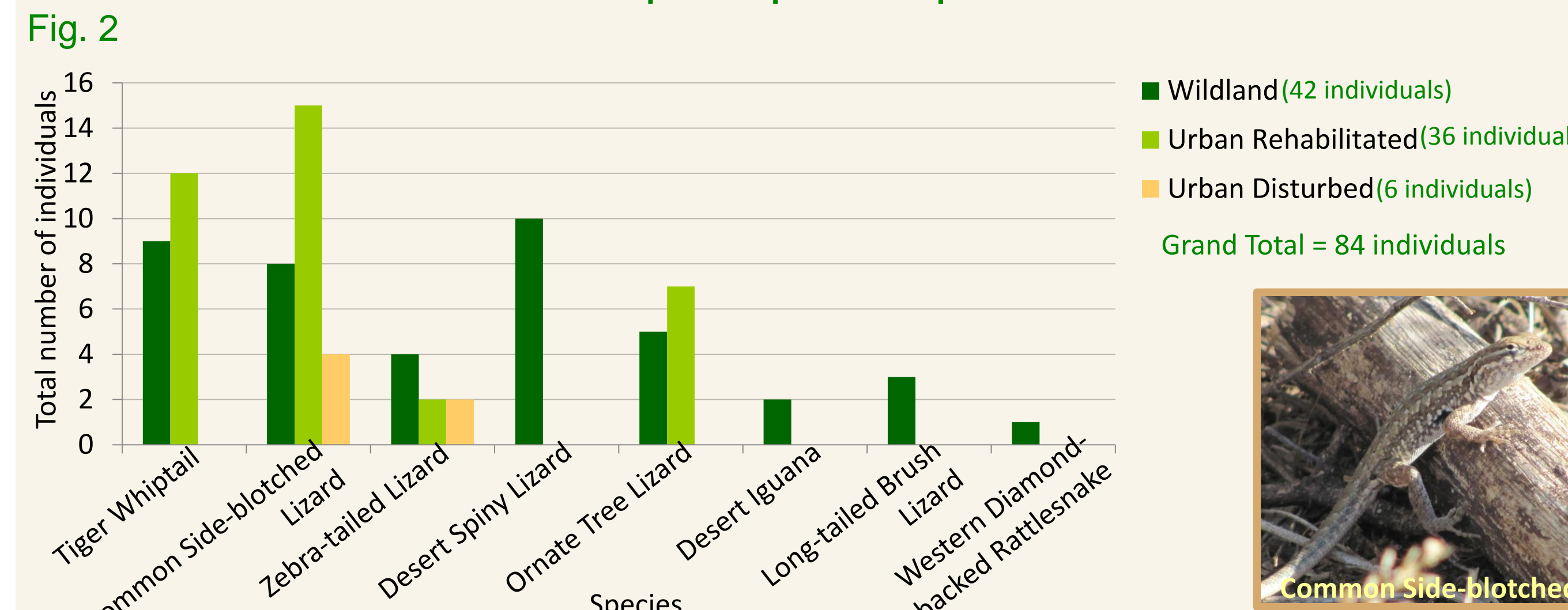
## Methods

- We performed herpetofauna visual surveys along 8 transects located at each reach (n=24).
- Transects were at least 150m apart, positioned 90° from stream. Three 20m x 10m segments per transect were surveyed (Fig.1).
- We collected various vegetation measurements with appropriate methods (i.e. line-intercept, etc.) to quantify microhabitat characteristics along transects.



## Results - Herpetofauna

### Total Number of Individual per Species per Reach



### Mean (SE) Maximum Abundance per Transect per Reach

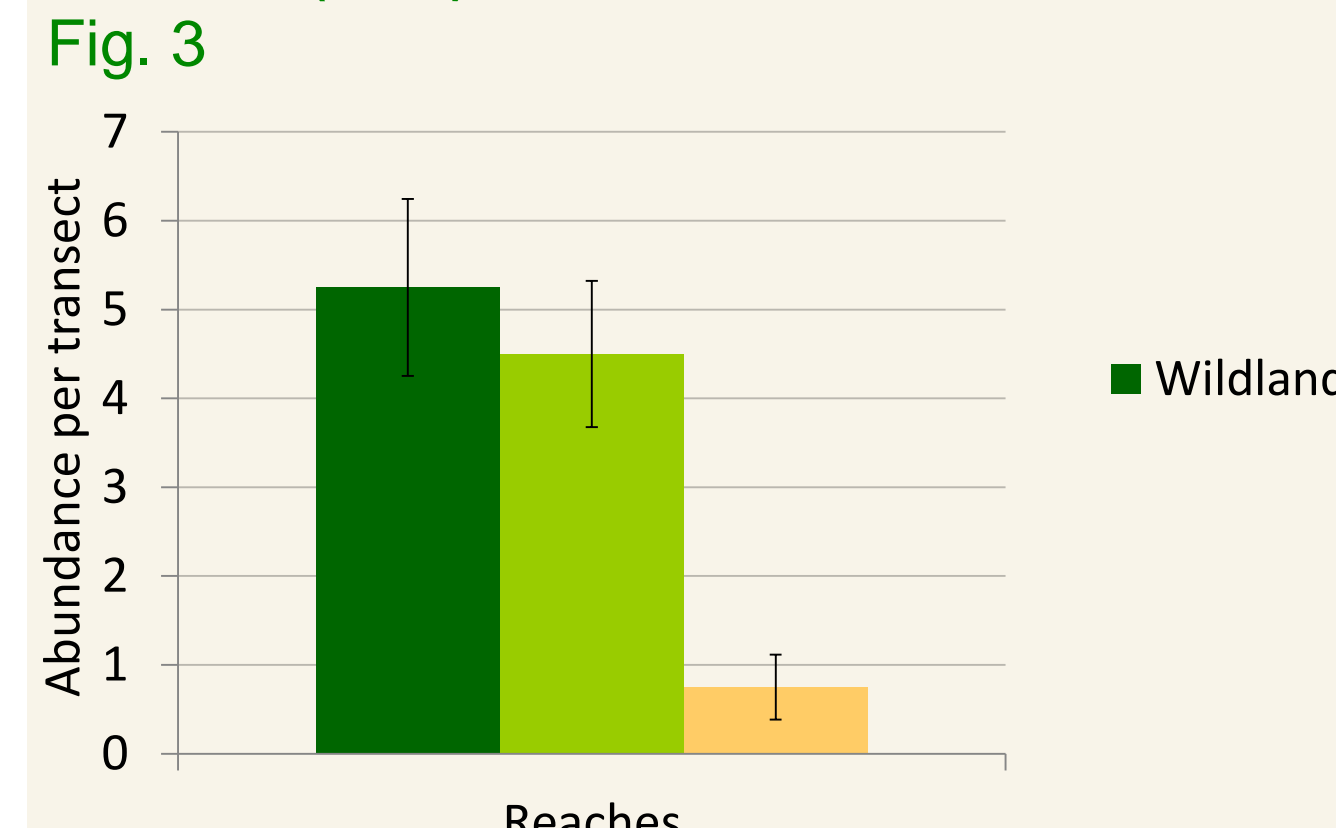


Fig. 3. Multiple Response Permutation Procedure (MRPP) and associated comparison test show that the abundance of the urban disturbed reach is approximately 6 times lower than the abundance of the two other reaches at p<0.001

### Mean (SE) Species Richness per Transect per Reach

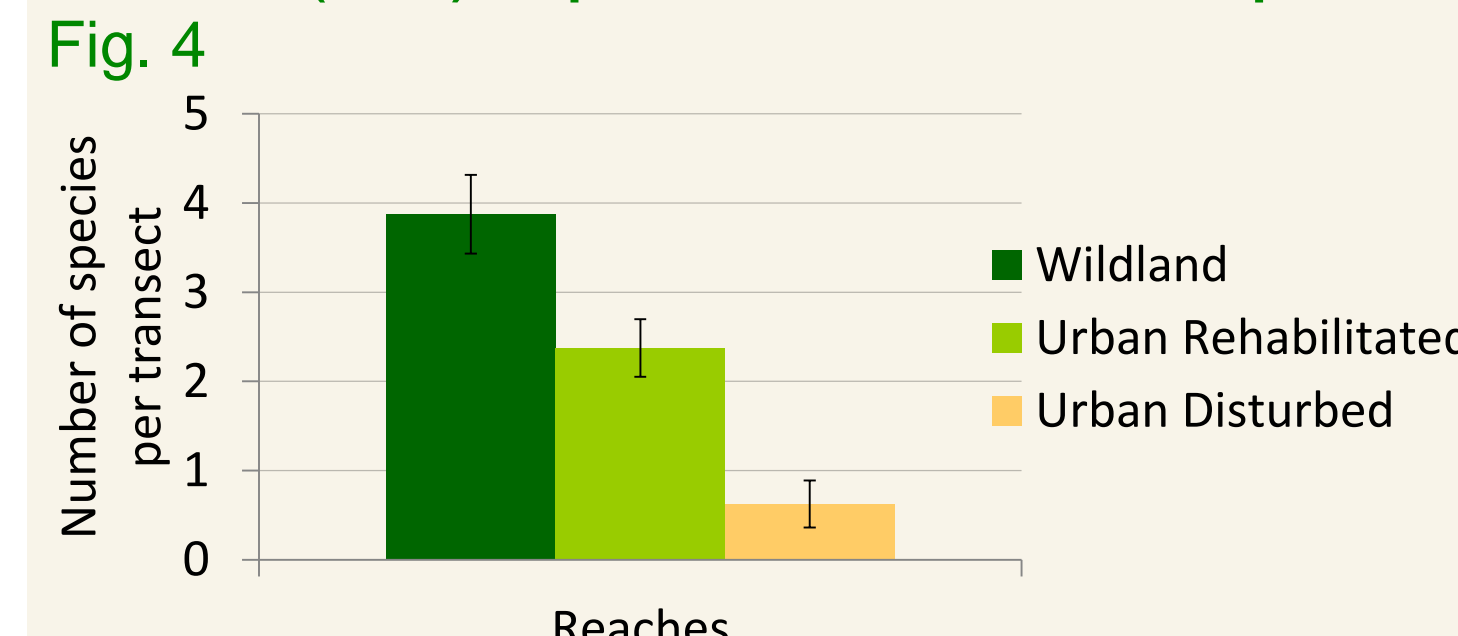


Fig. 4. MRPP and associated comparison test show that the species richness is significantly different between all reaches at p<0.05

### Diversity Ordering (Renyi)

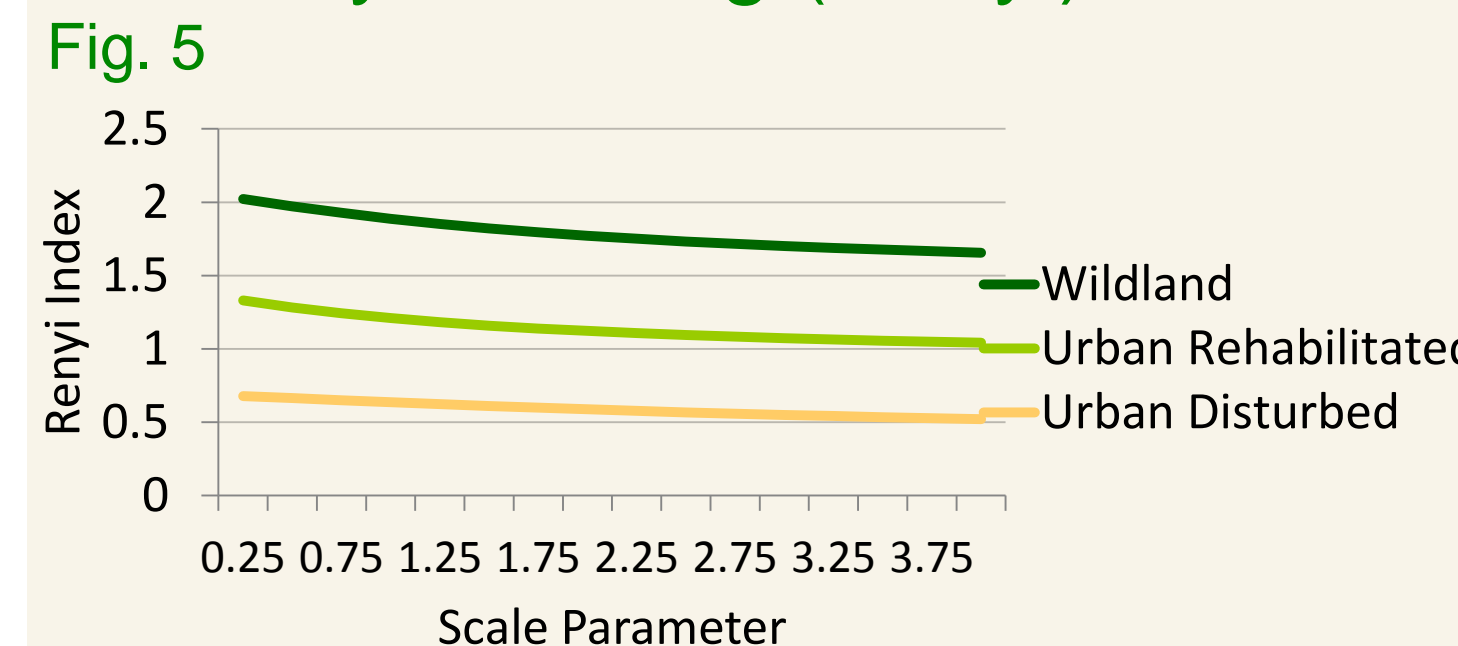


Fig. 5. Renyi index shows that the diversity of the three reaches rank consistently thus, are comparable

### Diversity Indices

- Shannon-Weiner and Simpson's diversity indices were computed and compared with software "Species Diversity and Richness 4.1.2"
- Species diversity of the wildland reach is significantly higher than the species diversity of the two urban reaches at p<0.05

## Results - Microhabitat

### Microhabitat Characteristics

Microhabitat Characteristics	Wildland	Urban Rehabilitated	Urban Disturbed	Factor # (correlation)
Bare ground (% cover)	46.0 (3.9)	67.9 (3.2)	95.6 (1.3)	1 (-)
Litter ground (% cover)	49.6 (3.8)	30.8 (2.8)	4.2 (1.4)	1 (+)
Tree shrub* (% cover)	47.8 (5.0)	16.6 (3.7)	0.1 (0.1)	1 (+)
Overstorey (% cover)	44.9 (6.6)	18.3 (5.6)	-	1 (+)
Woody debris per 10m	13.9 (1.5)	4.0 (2.5)	0.2 (0.1)	1 (+)
Vegetation species richness	11.9 (0.8)	12.0 (0.6)	2.4 (0.5)	1 (+)
Woody ground (% cover)	4.4 (0.8)	1.2 (0.8)	0.2 (0.2)	1 (+)
Shrub* (% cover)	5.6 (1.6)	3.2 (0.9)	-	1 (+)
Road/Trails (% cover)	3.3 (2.1)	12.6 (3.1)	0.4 (0.3)	2 (+)
Stem Prosopis/100m <sup>2</sup>	9.4 (3.2)	11.8 (3.2)	0.1 (0.1)	2 (+)
Litter depth (cm)	2.4 (0.4)	2.9 (1.3)	1.6 (0.8)	3 (+)
Subshrub* (% cover)	3.5 (1.0)	13.4 (3.0)	10.5 (2.0)	3 (+)
Tree* (%)	10.0 (4.1)	6.8 (3.7)	-	4 (+)
Burrows/200m <sup>2</sup>	28.6 (10.8)	6.1 (1.6)	2 (0.9)	4 (+)

\* Refers to "growth habit" as per USDA Plant Database website

Table 1. Mean (± SE) of microhabitat characteristics quantified along all transects per reach. Table also shows which factor the variables are loading on along with the variable correlation with the factor (positive or negative).

### Principal Component Analysis; 4 Factors Explain 84.7%

Variables	Description	Wildland	Urban Rehabilitated	Urban Disturbed	p value
Factor 1	Complex vegetative cover structure	1.13 (0.13)	0.02 (0.15)	-1.15 (0.02)	p < 0.00001
Factor 2	Mesquite and road/trail presence	-0.34 (0.37)	0.84 (0.33)	-0.50 (0.12)	p ≤ 0.02
Factor 3	Litter and <0.5m tall cover	-0.24 (0.17)	-0.008 (0.55)	0.24 (0.24)	p > 0.05
Factor 4	Presence of trees, lack of burrows	-0.40 (0.49)	0.55 (0.31)	-0.15 (0.03)	p > 0.05

Table 2. Mean (±SE) of Principal Component Analysis factors with eigen values ≥1. Those 4 factors explain 84.7% of the microhabitat characteristics variation. Table also shows significant differences for each factor between reaches. P values calculated with TukeyHSD at 95% confidence interval.

## Conclusion

### Preliminary Results Suggest:

- Rehabilitation may be beneficial for herpetofauna abundance.
- Urbanization may negatively influence herpetofauna diversity.

### Implication for Practice:

- Once ecological models are developed, we will be able to make suggestions as to how to rehabilitate an area to favor herpetofauna abundance and diversity.
- Herpetofauna abundance is important to favor as herps have an important ecological role in healthy ecosystems.
- Herpetofauna diversity is also important to favor as diverse ecosystems are typically more stable, more resistant, and/or more resilient to disturbance (Tilman & Downing 1994).

